



Kinetics and thermodynamics of lead (II) adsorption on lateritic nickel ores of Indian origin

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ABSTRACT

The lead adsorption from aqueous solution was studied in batch experiments using two typical Indian origin nickel lateritic ores having high (46.29%) and low iron content (28.56%) coded as NH and NL respectively. The adsorption was found to be strongly dependent on pH of the medium showing increase in uptake of Pb(II) from 11.0 to 53% and 8.2 to 44% for NH and NL samples respectively with the increase in pH in the range of 2.0–5.23. The time data generated at different temperatures for both the samples fitted well to second-order kinetic model and Elovich equation. The later is indicative of a chemisorption process. The +ve ΔH° values (8.90 and 10.29 kJ mol⁻¹ for NH and NL samples) support the endothermic nature of adsorption. The +ve ΔS° values (28.56 and 29.40 kJ mol⁻¹ K⁻¹ for NH and NL respectively) suggest that the adsorption occurs with internal structural changes. The activation energy was estimated to be 7.6 and 3.12 kJ mol⁻¹ for NH and NL respectively. The thermodynamic activation parameters were also evaluated using Eyring equation. The loading capacities of NH and NL were estimated to be 44.4 and 28.45 mg g⁻¹ respectively under the experimental conditions: adsorbent concentration 2 g l⁻¹, time 30 min, temperature 308 K and pH 5.23. Data fitted well to Langmuir and Freundlich isotherm models for NH while in case of NL only Langmuir isotherm showed good fit. Due to high loading capacities and favorable kinetics, these materials can be utilized for Pb(II) removal from aqueous solutions.

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1. Introduction

Pb(II) is one of the toxic heavy metal ions notably present in soil environments through atmospheric dusts and solid wastes from a variety of active and inactive sources [1]. Its presence in the environment as a result of extensive and wide applications of mining, chemical, electroplating, petroleum refining, paper and pulp industries poses serious health hazards to living organisms. The permissible limit for Pb(II) in wastewater given by the Environmental Protection Agency is 0.05 mg l⁻¹ [2] and that of the Bureau of Indian Standards (BIS) is 0.1 mg l⁻¹ [3]. Therefore, it is needed to reduce the amount of this ion particularly in wastewater streams of hydrometallurgical and other industries. Filtration, adsorption, reverse osmosis, solvent extraction, electro-kinetic and membrane separation techniques are used for removal of heavy metals from aqueous solutions [4–18]. Among them adsorption has been shown to be an economically feasible alternative method for removing heavy metals from wastewater and water supplies [11–18]. Low cost adsorbents are generally preferred for the removal of impurities from aqueous solutions/contaminated water. Therefore, the search for novel materials with high efficiency to remove heavy metals

from water and industrial effluents is of environmental and industrial interest. Waste products such as waste from bauxite refining (red mud), fly ash [19], blast furnace slag [20], manganese nodule leach residue [21] have been studied as adsorbents for the removal of Pb(II) ions. However, no work has been reported on the use of low and off grade ores for adsorption purposes. The objective of this study is to investigate and explore the feasibility of using nickel lateritic ores of Orissa, India for adsorption of Pb(II). The iron content in the Indian ore varies between 20 and 50% whereas the nickel content varies between 0.5 and 1.2%. These ores have not been commercially exploited for nickel extraction till now due to low nickel content. However, considerable efforts have been made at Institute of Minerals and Materials Technology to beneficiate these ores for upgrading nickel content and for developing acid/ammoniacal based processes [22–25] but still these ores are lying unutilized. With a view to find an alternative application of these ores for effluent and wastewater treatment, the present work on removal of Pb(II) from aqueous solution was taken up.

2. Experimental

2.1. Materials

Bulk samples of typical low and high iron containing nickel laterites samples were collected from Sukinda, Orissa, India. The

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Nomenclature

| | |
|---------------------|--|
| q_e | equilibrium concentration on adsorbent, mg g^{-1} |
| r^2 | linear regression coefficient |
| t | time, min |
| T | temperature, K |
| q_t | the amount of Pb(II) adsorbed per unit mass of the adsorbent at time t , mg g^{-1} |
| k_1 | adsorption rate constant for pseudo-first-order kinetic equation, t^{-1} |
| k_2 | adsorption rate constant for pseudo-second-order kinetic equation, $\text{g mg}^{-1} \text{min}$ |
| E_a | activation energy, kJ mol^{-1} |
| R | gas constant |
| k_0 | Arrhenius factor |
| k_B | Boltzmann's constant |
| h | Planck's constant |
| ΔG° | free energy of reaction, kJ mol^{-1} |
| ΔH° | enthalpy of reaction, kJ mol^{-1} |
| ΔS° | entropy of reaction, $\text{J mol}^{-1} \text{K}^{-1}$ |
| ΔG^\ddagger | free energy of activation, kJ mol^{-1} |
| ΔH^\ddagger | enthalpy of activation, kJ mol^{-1} |
| ΔS^\ddagger | entropy of activation, $\text{kJ mol}^{-1} \text{K}^{-1}$ |

samples were ground and sieved to obtain 100% <100 μm (–150 Mesh BSS, British Standard Sieves). $\text{Pb}(\text{NO}_3)_2 \cdot 4\text{H}_2\text{O}$, H_3PO_4 , tin granules, mercurous chloride used in the present work were of MERCK, India, while NaOH, HCl, H_2SO_4 were of BDH grade and, BDAS indicator was of s-d-fine chemicals.

2.2. Characterization of adsorbents

A weighed amount of sample was subjected to tri-acid digestion for wet chemical analysis. Iron was analysed volumetrically [26]. The pH_{pzc} of prepared sample was determined following Balistrieri and Murray method [27]. The X-ray diffraction (XRD) measurements were done over a range of 20–70° using Phillips Powder Diffractometer Model PW1830 with $\text{Co K}\alpha$ radiation at a scan speed of 1.2°/min. Surface area was measured with Quantasorb 1750 instrument.

2.3. Adsorption experiments

The Pb(II) adsorption experiments were carried out in an horizontal shaker having provisions for temperature and agitation control. After proper dilutions Pb(II) analysis was done by Atomic Absorption Spectrophotometer (PerkinElmer Model 2380). For each experiment 50 ml of Pb(II) solution of desired concentration and pH was taken in 100 ml stoppered conical flask and a weighed amount of adsorbent was added to it. The contents were stirred at 160 rpm (revolutions per minute) for desired period of time at constant temperature. The kinetic of adsorption was studied at various temperatures while keeping the adsorbent dose as 2 g l^{-1} . For fitting adsorption isotherm the data was generated at different initial

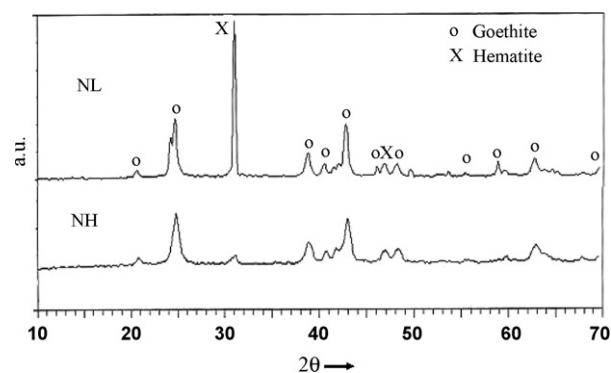


Fig. 1. XRD patterns of NH and NL samples.

concentrations of Pb(II) (50–500 mg l^{-1}) while keeping the adsorbent dose fixed (2 g l^{-1}). In order to test the reproducibility, some of the experiments were carried in duplicate and the reproducibility was found to be within $\pm 2\%$.

3. Results and discussions

3.1. Chemical analysis and characterization

The physico-chemical analyses of two laterite samples are given in Table 1. For convenience the high iron containing sample has been coded as NH and the low iron containing one as NL. The XRD pattern (Fig. 1) shows goethite ($\alpha\text{-FeOOH}$) (JCPDS, 1980, File No 29-713) as the major crystalline phase with 100% RI (relative intensity) peak at d value of 4.2 Å for NH sample. Some minor peaks corresponding to hematite ($\alpha\text{-Fe}_2\text{O}_3$) (JCPDS, 1980, File No. 33-0664) are also observed. In case of NL sample, the XRD pattern shows both hematite and goethite as the major phases though the 100% RI peak at d value of 2.7 Å corresponds to hematite.

3.2. Effect of contact time at various pH

Effect of pH was studied in the range of 2–5.23 at 308 K with rest of conditions as: adsorbate concentration 100 mg l^{-1} and adsorbent concentration 2 g l^{-1} . The results at various pH are shown in Figs. 2 and 3 for NH and NL samples. The pH was not further increased as on keeping the 100 mg l^{-1} Pb(II) containing solution at a pH of 5.5, about 7% Pb(II) precipitated. With the increase in pH from 2 to 5.23, % Pb(II) adsorption increased from 11 to 53% for NH (Fig. 2) and from 8.2 to 44% for NL samples (Fig. 3). This could be due to increase in negative surface charge with increase in pH thereby facilitating cations adsorption. Furthermore, the proportion of hydrated ions increases with pH and these may have higher affinity as compared to the unhydrated Pb(II) ions [28]. Therefore, both these effects would contribute to higher cation adsorption with increase in pH. The results given in Figs. 2 and 3 show that equilibrium is attained in 60 min and no significant increase in adsorption is observed by further increasing the contact time. For rest of the experiments contact time was kept as 1 h.

Table 1
Physico-chemical properties of nickel laterites.

| Sample | Chemical analysis, % metal ^a | | | | | | Surface area ($\text{m}^2 \text{g}^{-1}$) | pH_{pzc} |
|--------|---|------|------|------|------|------|---|--------------------------|
| | Fe | Ni | Cu | Co | Mn | Cr | | |
| NH | 46.29 | 1.15 | 0.04 | 0.07 | 1.06 | 2.24 | 74.67 | 6.95 |
| NL | 28.56 | 0.71 | 0.02 | 0.05 | 0.27 | 1.7 | 68.39 | 6.70 |

^a All the metals are present as oxides/hydroxides. The iron is present as goethite and hematite. The balance will be acid insolubles which are 28.5% for NH and 55.3% for NL.

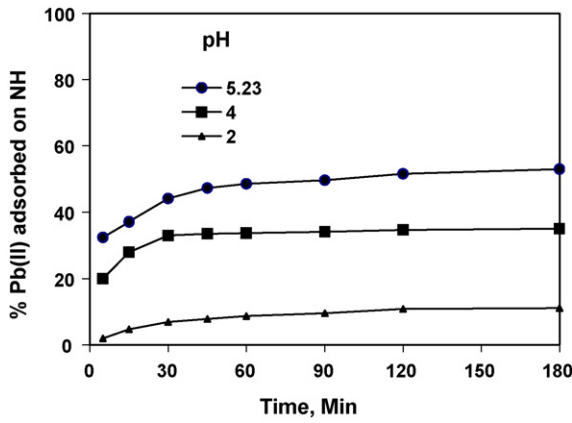


Fig. 2. Effect of contact time at various pH on adsorption of Pb(II) on NH: adsorbate concentration 100 mg l^{-1} , adsorbent concentration 2 g l^{-1} , temperature 308 K .

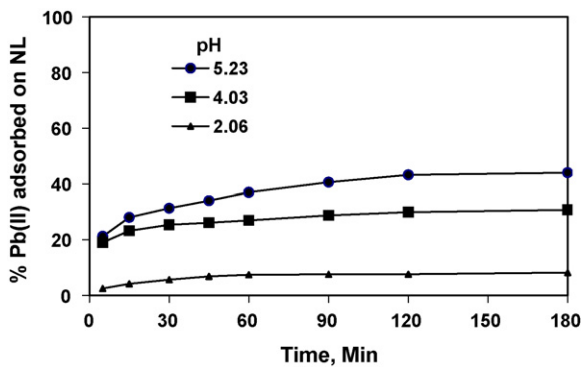


Fig. 3. Effect of contact time at various pH on adsorption of Pb(II) on NL: adsorbate concentration 100 mg l^{-1} , adsorbent concentration 2 g l^{-1} , temperature 308 K .

3.3. Effect of temperature

Effect of temperature was studied in the range of $308\text{--}338 \text{ K}$ at a pH of 5.23 while rest of the conditions were kept as: Pb(II) 100 mg l^{-1} and adsorbent concentration 2 g l^{-1} . The adsorption data was generated for 60 min (Figs. 4 and 5). An increase in the temperature from 308 to 338 K led to an increase in % adsorption from 48.57 to 56% for NH and from 38.33 to 46.12% for NL samples. This is characteristic of a chemical reaction or bond being involved in the adsorption process. The increase in adsorption could be due to changes in pore size or an increase in kinetic energy of

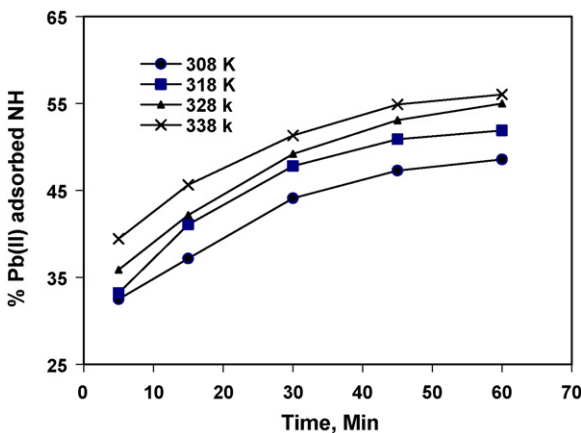


Fig. 4. Effect of temperature on adsorption of Pb(II) on NH: adsorbate concentration 100 mg l^{-1} , adsorbent concentration 2 g l^{-1} and pH 5.23 .

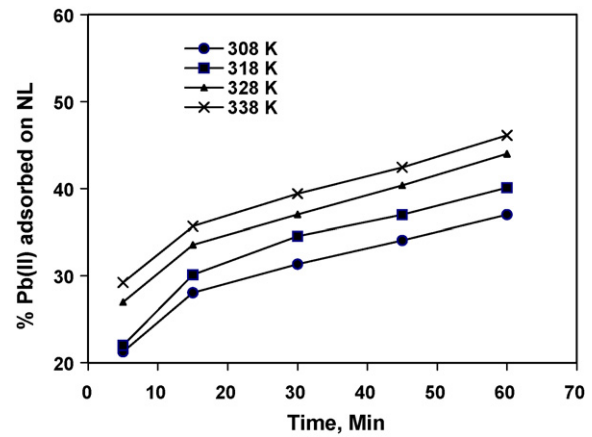


Fig. 5. Effect of temperature on adsorption of Pb(II) on NL: adsorbate concentration 100 mg l^{-1} , adsorbent concentration 2 g l^{-1} and pH 5.23 .

the adsorbents. Such observations for adsorption of lead on iron oxide containing surfaces have been reported earlier [21,29,30]. The increase in % Pb(II) adsorption with increase in temperature suggests endothermic nature of adsorption which is attributable unequivocally to chemisorption [31].

3.4. Thermodynamic parameters

The thermodynamic quantities such as change in free energy (ΔG°), change in enthalpy of adsorption (ΔH°) and change in entropy (ΔS°) may give an insight about the nature and mechanism of adsorption process.

The values of ΔH° and ΔS° were determined from the Van't Hoff equation as given below.

$$\log K_c = \frac{\Delta S^\circ}{2.303R} - \frac{\Delta H^\circ}{2.303RT} \quad (1)$$

where R is the gas constant, T is the temperature (in Kelvin) and K_c is distribution coefficient and is determined as:

$$K_c = \frac{C_A}{C_e} \quad (2)$$

where C_e is the equilibrium concentration in solution (mg l^{-1}) and C_A is the adsorbed amount of adsorbate at equilibrium (mg l^{-1}) [32].

From the slopes and intercepts of the Van't Hoff plots (Fig. 6) ΔH° and ΔS° were calculated and are given in Table 2. Both ΔH° and ΔS° values are found to be positive suggesting a chemisorption process. During chemisorption, a certain amount of destructuring (i.e., desolvation) may be involved, creating a net positive entropy change and a slightly positive enthalpy of adsorption [33]. The positive value of adsorption entropy indicates the irreversible adsorption

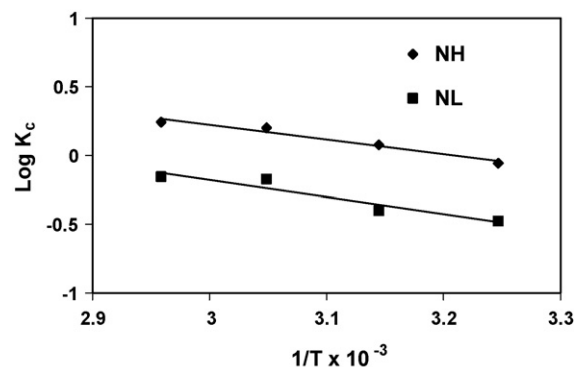


Fig. 6. Van't Hoff plots (data corresponds to Figs. 4 and 5).

Table 2

Thermodynamic and activation parameters of Pb(II) adsorption on NH and NL samples.

| | NH | | | | NL | | | |
|---|--------|--------|--------|--------|--------|-------|-------|-------|
| | 308 | 318 | 328 | 338 | 308 | 318 | 328 | 338 |
| Temperature (K) | 308 | 318 | 328 | 338 | 308 | 318 | 328 | 338 |
| ΔG° (kJ mol ⁻¹) | 0.103 | -0.182 | -0.467 | -0.753 | 1.244 | 0.95 | 0.656 | 0.362 |
| ΔH° (kJ mol ⁻¹ K ⁻¹) | 8.90 | | | | 10.29 | | | |
| ΔS° (J mol ⁻¹) | 28.56 | | | | 29.40 | | | |
| E_a | 7.67 | | | | 3.12 | | | |
| ΔG^\ddagger (kJ mol ⁻¹) | 30.00 | 30.82 | 31.63 | 32.44 | 27.81 | 28.67 | 29.54 | 30.41 |
| ΔH^\ddagger (kJ mol ⁻¹ K ⁻¹) | 4.99 | | | | 1.1 | | | |
| ΔS^\ddagger (J mol ⁻¹) | -81.23 | | | | -86.71 | | | |

process, increased randomness at the solid–solution interface and stability of adsorption [33,34].

The free energy (ΔG°) values are calculated using the following equation:

$$\Delta G^\circ = \Delta H^\circ - T\Delta S^\circ \quad (3)$$

The free energy (ΔG°) values are given in Table 2. The ΔG° values are negative for NH except at 308 K. When the temperature increased from 318 to 338 K, the magnitude of free energy change (ΔG°) shifted to higher negative values. The ΔG° values for NL samples were +ve in the studied range of temperature.

3.5. Kinetics of adsorption

The experimental data of Figs. 4 and 5 were tested for best fits for pseudo-first-order Eq. (4) [35], the pseudo-second-order Eq. (5) [36] and Elovich Eq. (6) [37]. These equations are given below:

$$\text{1st order kinetics } \ln(q_e - q_t) = \ln q_e - k_1 t \quad (4)$$

$$\text{2nd order kinetics } \frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \left(\frac{1}{q_e}\right) t \quad (5)$$

$$\text{Elovich equation } q_t = \beta \ln(\alpha\beta) + \beta \ln t \quad (6)$$

From this simple form of Elovich equation, the parameters are estimated without using the origin ($q=0$; $t=0$).

Where q_e and q_t are the amounts of metal ion adsorbed per unit weight of adsorbent at equilibrium and at any time t , respectively (mg g^{-1}) and k_1 is the rate constant of pseudo-first-order sorption (min^{-1}). k_2 is the rate constant of pseudo-second-order adsorption ($\text{g mg}^{-1} \text{min}$). α is the initial adsorption rate of Elovich equation ($\text{mg g}^{-1} \text{min}^{-1}$), and the parameter β is related to the extent of surface coverage (mg g^{-1}) and activation energy for chemisorption [38]. The constants can be obtained from the slope and intercept of a straight line of q_t vs $\ln t$.

The Pb(II) adsorption data given in Figs. 4 and 5 were treated to evaluate the rate constants at different temperatures for NH and NL samples. The data was first treated by first-order rate expression. The first-order rate constant, k_1 , the regression coefficient r^2 and theoretical and experimental equilibrium adsorption capacity, q_e are given in Table 3. The $\log(q_e - q_t)$ vs time (t) plot (not shown) were linear with r^2 values varying from 0.972 to 0.987 for NH and from 0.944 to 0.993 for NL sample. However, the theoretical and experimental equilibrium adsorption capacities, q_e obtained from these plots varied widely. This indicates that this equation cannot provide an accurate fit of the experimental data. The data was then treated to pseudo-second-order kinetic plot (Eq. (5)). The t/q_t vs t plots are shown in Figs. 7 and 8 for NH and NL respectively. All the parameters obtained from these graphs are given in Table 3. The adsorption data for both the materials are found to fit well to pseudo-second-order kinetic as (i) the regression coefficients, r^2 for both NH and NL plots were ≥ 0.986 , (ii) the theoretical as well as experimental q_e values matched very well ($\Delta q_e < 5\%$). An increase in temperature resulted in increase of q_e which is typical to chemisorption nature of adsorption [39,40]. Further the

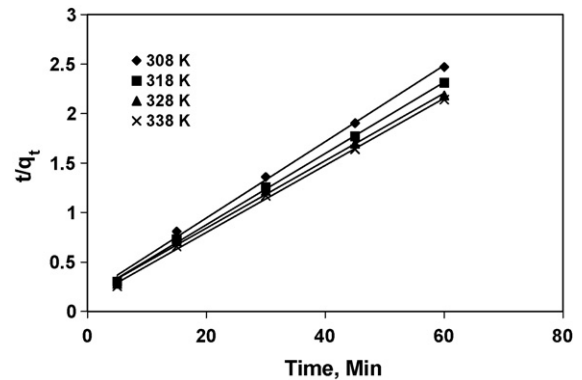


Fig. 7. Pseudo-second-order kinetic plots for NH (data corresponds to Fig. 4).

data was fitted to Elovich equation which is of general application to chemisorption kinetics. Many researchers [38–41] applied this kinetic model satisfactorily to the chemisorption processes. It is often valid for systems in which the adsorbing surface is heterogeneous. The Elovich plots (q_t vs $\ln t$) are given in Figs. 9 and 10. The regression coefficients determined from these plots show good linearity for NH ($r^2 = 0.972$ – 0.993) and NL samples ($r^2 = 0.953$ – 0.986). The values of the coefficient α and β are also given in Table 3. With the increase in temperature, the constant α and β increased showing that both the rate of chemisorption and the available adsorption surface would increase. In conclusion, the pseudo-second-order model provides better correlation of the sorption data than the pseudo-first-order model, thus suggesting that the rate limiting step may be chemical sorption rather than diffusion.

3.6. Activation parameters

The values of the rate constant k_2 at different temperatures listed in Table 3 were applied to estimate the activation energy of the adsorption of Pb(II) onto NH and NL samples by the Arrhenius

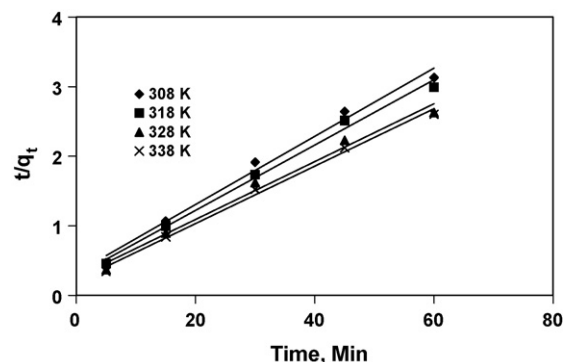


Fig. 8. Pseudo-second-order kinetic plots for NL (data corresponds to Fig. 5).

Table 3
Rate coefficients for adsorption of Pb(II) on NH and NL.

| | NH | | | | NL | | | |
|---|-------|-------|-------|-------|-------|-------|-------|-------|
| Temperature (K) | 308 | 318 | 328 | 338 | 308 | 318 | 328 | 338 |
| q_{exp} (mg g ⁻¹) | 24.28 | 25.95 | 27.49 | 28.02 | 19.16 | 20.05 | 22.83 | 23.06 |
| Pseudo-first-order | | | | | | | | |
| $k_1 \times 10^{-2}$ (min ⁻¹) | 6.4 | 7.2 | 5.75 | 6.48 | 3.33 | 3.59 | 3.04 | 3.68 |
| q_e (mg g ⁻¹) | 12.99 | 15.01 | 14.41 | 13.21 | 9.37 | 9.45 | 9.76 | 10.35 |
| r^2 | 0.977 | 0.987 | 0.985 | 0.972 | 0.984 | 0.944 | 0.988 | 0.993 |
| Pseudo-second-order | | | | | | | | |
| $k_2 \times 10^{-3}$ (g mg ⁻¹ min) | 3.14 | 3.39 | 3.73 | 4.09 | 7.3 | 7.75 | 7.95 | 8.25 |
| q_e (mg g ⁻¹) | 25.9 | 30.67 | 31.64 | 32.36 | 20.48 | 21.32 | 24.03 | 24.21 |
| r^2 | 0.997 | 0.995 | 0.997 | 0.998 | 0.988 | 0.992 | 0.986 | 0.993 |
| Elovich model | | | | | | | | |
| α (g min ² mg ⁻¹) | 6.02 | 5.33 | 5.61 | 6.24 | 1.64 | 1.87 | 2.57 | 5.34 |
| β (mg g ⁻¹ min ⁻¹) | 10.31 | 11.93 | 12.31 | 13.13 | 5.31 | 5.53 | 7.55 | 9.19 |
| r^2 | 0.972 | 0.993 | 0.985 | 0.991 | 0.973 | 0.983 | 0.953 | 0.986 |

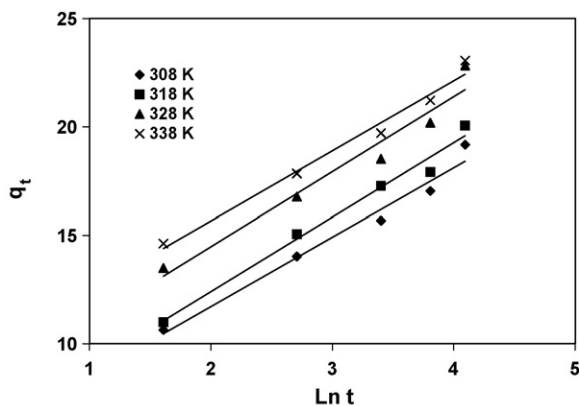


Fig. 9. Elovich kinetic plots for NH (data corresponds to Fig. 4).

equation [42] as follows:

$$\ln k_2 = -\frac{E_a}{RT} + \ln k_0 \quad (7)$$

where k_0 is the Arrhenius factor and R is the gas constant.

A straight line is obtained by plotting of the logarithm of the rate constant against the reciprocal of the absolute temperature (Fig. 11) and the activation energy E_a are listed in Table 2. The apparent activation energy of adsorption was found to be 7.6 and 3.12 kJ mol⁻¹ for NH and NL, respectively (Table 2). These relatively low activation energy values obtained indicate that the adsorption of cations on the adsorbent surface is rapid; the rate limiting step being the migration of the cations to the adsorption site [43,44].

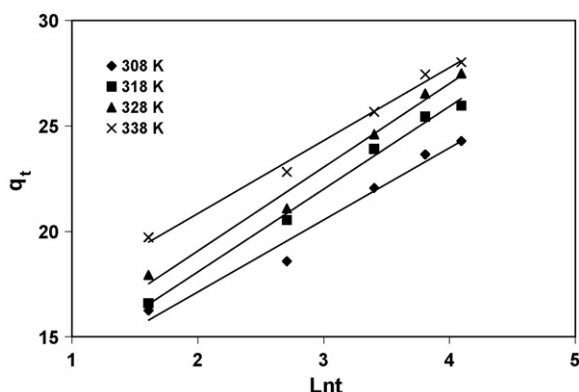


Fig. 10. Elovich kinetic plots for NL (data corresponds to Fig. 5).

To get an insight whether the adsorption process follows an activated complex, the enthalpy (ΔH^\ddagger), entropy (ΔS^\ddagger) of activation were calculated using Eyring equation (Eq. (8)) [45]. Free energy (ΔG^\ddagger) of activation was calculated using Eq. (9).

$$\frac{\ln k_2}{T} = \left(\frac{\ln k_B}{h_p} + \frac{\Delta S^\ddagger}{R} \right) - \frac{\Delta H^\ddagger}{RT} \quad (8)$$

$$\Delta S^\ddagger = \frac{\Delta G^\ddagger - \Delta H^\ddagger}{T} \quad (9)$$

where k_B is Boltzmann's constant and h_p is Planck's constant, other parameters are defined earlier. From the slope and intercept of the graph $\ln k_2/T$ vs $1/T$ (Fig. 12), the enthalpy (ΔH^\ddagger) and entropy (ΔS^\ddagger) of activation were calculated. The enthalpy of activation, ΔH^\ddagger values were 4.99 and 11.0 kJ mol⁻¹ (Table 2) for NH and NL respectively. The magnitude and sign of ΔS^\ddagger gives an indication whether the adsorption reaction is an associative or dissociative mechanism [46,47]. The negative value of ΔS^\ddagger suggests that Pb(II) adsorption onto NH and NL is an associative mechanism.

3.7. Effect of Pb(II) concentration

With a view to generate adsorption isotherms, the initial concentration of Pb(II) was varied from 50 to 500 mg l⁻¹ at a constant adsorbent concentration of 2 g l⁻¹, at pH of 5.23 and contact time of 60 min. The extent of adsorption decreased as expected from 69.77 to 17.77% for NH and from 46.72 to 11.38% for NL. The adsorption

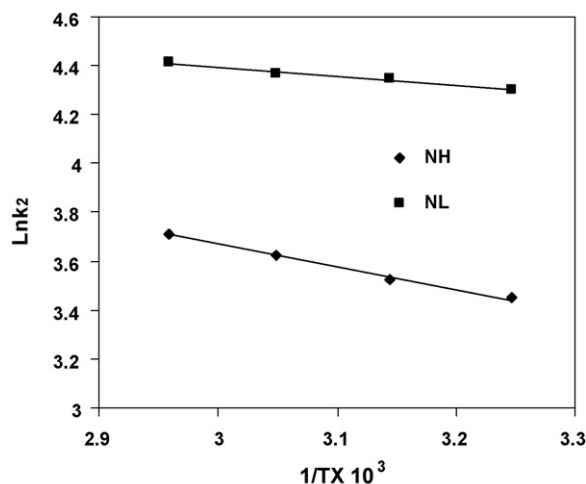


Fig. 11. Arrhenius plots: Pb(II) concentration 100 mg l⁻¹, adsorbent concentration 2 g l⁻¹ and pH 5.23.

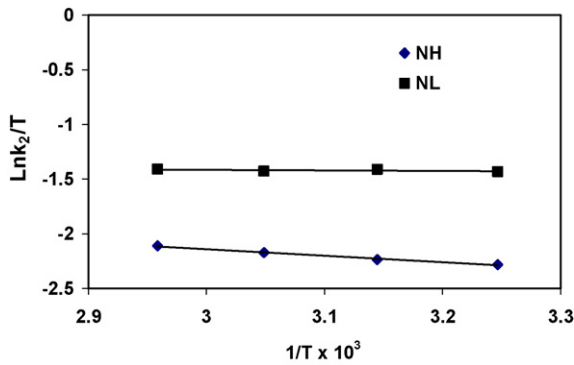


Fig. 12. Eyring plots: Pb(II) concentration 100 mg l⁻¹, adsorbent concentration 2 g l⁻¹ and pH 5.23.

isotherms showing q_e vs C_e plots are given in Fig. 13. Under the experimental conditions the maximum loading capacities of NH and NL are found to be 44.4 and 28.45 mg g⁻¹ respectively. Higher loading capacity of NH may be attributed to the high iron content of the sample.

3.8. Adsorption isotherms

Most widely tested isotherms for adsorption process are Langmuir and Freundlich models [48,49] which are given by Eqs. (10) and (11):

$$\text{Langmuir } \frac{C_e}{q_e} = \frac{1}{bq_m} + \left(\frac{1}{q_m}\right) C_e \quad (10)$$

$$\text{Freundlich } q_e = K_f C_e^n \quad (11)$$

where q_m , is the monolayer adsorption capacity (mg g⁻¹) and 'b', the constant related to the energy of adsorption (l g⁻¹), K_f indicates adsorption capacity (mg g⁻¹) and 'n' an empirical parameter related to the intensity of adsorption, which varies with the heterogeneity of the adsorbent. For "n" values in the range 1–10 (0.1 < 1/n < 1), adsorption is favorable. The greater the values of 'n' better is the favorability of adsorption.

According to Hall et al. [50], the essential features of the Langmuir isotherm can be expressed in terms of a dimensionless constant separation factor or equilibrium parameter R_L , expressed by Eq. (9).

$$\text{Separation factor } R_L = \frac{1}{1 + bC_e} \quad (12)$$

The isotherm is (i) unfavorable when $R_L > 1$, (ii) linear when $R_L = 1$, (iii) favorable when $R_L < 1$, and (iv) irreversible when $R_L = 0$. The value of R_L lie between 0 and 1 for favorable adsorption process

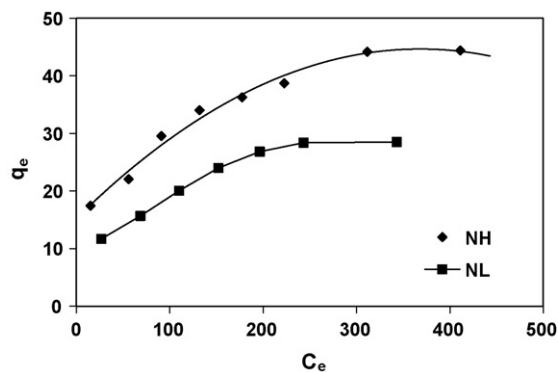


Fig. 13. Adsorption isotherms for NH and NL samples (q_e vs C_e): adsorbent concentration 2 g l⁻¹, pH 5.23 and temperature 308 K.

Table 4

Langmuir and Freundlich parameters for adsorption of Pb(II) on NH and NL.

| Adsorbents | Langmuir coefficients | | | Freundlich coefficients | | | |
|------------|-----------------------------|--------------------------|-------|-------------------------|-------|------|-------|
| | q_m (mg g ⁻¹) | b (l g ⁻¹) | r^2 | R_L | K_f | n | r^2 |
| NH | 50.25 | 0.017 | 0.987 | 0.63 | 4.62 | 3.28 | 0.970 |
| NL | 33 | 0.017 | 0.991 | 0.53 | 5.18 | 2.84 | 0.940 |

[51]. The experimental data corresponding to Fig. 13 was treated for Langmuir and Freundlich models. The values of the adsorption coefficients obtained from the plots (not shown) are given in Table 4. The values of the Langmuir equilibrium coefficient, 'b', were found to be same (0.017 l g⁻¹) for both the samples. The Langmuir monolayer capacity, q_m , values were found to be 50.25 mg g⁻¹ for NH and 33 mg g⁻¹ for NL sample which are quite close to the experimental values. The separation factor R_L values of 0.63 and 0.53 for NH and NL respectively, indicate that the metal ions prefer to be in the bound state with surfaces [52]. The Freundlich isotherms (plot not shown) also yield good linear plots (r^2 were 0.97 and 0.94 for NH and NL). The Freundlich empirical coefficient, 'n' is found to be greater than one which indicates, adsorption intensity is high and adsorption occur through ion exchange process [53]. The fact that the adsorption of Pb(II) on NH and NL samples follow the Langmuir isotherm models indicates that a monolayer coverage is formed on the surface of the adsorbent [54]. Even though the Freundlich plot does yield a linear curve, the curve seems to bend at higher values of C_e , as the adsorption approaches a maximum value, confirming complete monolayer coverage of the adsorbent particles [34]. From these results the monolayer formation during adsorption of Pb(II) is confirmed.

3.9. Conclusions

Pb(II) adsorption on high and low iron containing laterite samples (NH and NL samples) increased with an increase in pH (2–5.23) and temperature (308–338 K). The adsorption kinetics followed pseudo-second-order model indicating towards chemisorption. Maximum Pb(II) adsorption at pH 5.23, temperature 308 K, initial Pb(II) concentration 400 mg l⁻¹ were 44.4 and 28.4 mg Pb(II) g⁻¹ of NH and NL samples respectively. The adsorption process was endothermic ($\Delta H = 8.90$ and 10.29 kJ mol⁻¹ for NH and NL samples) accompanied by positive entropy ($\Delta S = 28.56$ and 29.40 kJ mol⁻¹ deg⁻¹). The activation energy values (7.6 and 3.12 kJ mol⁻¹ for NH and NL) indicate the adsorption occur via an activated chemisorption step with migration of cations to the adsorption site being the rate limiting step. The negative value of entropy of activation suggests that Pb(II) adsorption onto NH and NL is an associative mechanism. The adsorption data fitted well to both the Langmuir and Freundlich isotherms models. The isothermic data indicated monolayer coverage of Pb(II) on high and low iron containing laterites.

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